Short Communication



Optimisation of pH, Temperature and Carbon Nitrogen Ratio for the Degradation of *m*-Chlorophenol by *Pseudomonas putida* CP1

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Aromatic pollutants like *m*-chlorophenol is toxic to the environment and chlorophenol containing a *meta*chlorine are more persistent under aerobic conditions than compounds lacking a chlorine substituent in positions *meta* to hydroxyl group. Therefore, it should be removed effectively from the environment. In order to increase the degradative activity, the optimum conditions for *m*-chlorophenol degradation by *Pseudomonas putida* CP1, some physicochemical conditions like pH, temperature and carbon nitrogen ratio for the growth and degradation of most persistent monochlorophenol, *m*-chlorophenol by the organism was optimised. The pH optimum for *m*-chlorophenol degradation by the bacterium was between pH 6.5 and 7.0 and the temperature optimum was 30° C for removal activity. Carbon : nitrogen (C:N) ratio of 3:1 was found best for effective removal of chemical oxygen demand (COD) and *m*-chlorophenol by the bacterium.

Keywords: m-Chlorophenol degradation, Pseudomonas putida CP1, Chemical oxygen demand (COD)

Halogenated aromatics, particularly chlorinated aromatics, are produced in vast quantities due to their numerous applications such as herbicides, insecticides, fungicides, solvents, hydraulic and heat transfer fluids, plasticizers, and intermediates for chemical synthesis. Because of their toxicity, bio-concentration, and persistence, the ubiquitous distribution of the halogenated compounds in the biosphere has caused public concern over the possible effects on the quality of life¹. Though some chlorinated aromatic compounds are biodegradable, they are often recalcitrant and natural purification of contaminated sites seems slow because microorganisms able to degrade the contaminant are missing or that the environmental conditions e.g. temperature, redox potential, pH, and concentration of contaminant are such that degradation is not promoted. Environmental pH is the most important factor affecting chlorophenol adsorption and mobility. Soil organic content is another important factor affecting chlorophenol mobility².

Chlorophenols are much more environmentally stable than the parent unsubstituted phenol. As the number of chlorine substituents increases the rate of aerobic decomposition decreases³. Compounds containing a *meta*-chlorine are more persistent under aerobic conditions than compounds lacking a chlorine substituent in positions *meta* to hydroxyl group⁴⁻⁵. Persistence of chlorophenols in the environment depend on the presence of microbial populations and environmental parameters such as pH, temperature, aeration rate, available nutrients, the

absence or presence of inhibitory co-pollutants, and the absence or presence of substances changing the electron flow in the system. Chlorinated phenols may be removed from a water body via volatilisation, photo-degradation, adsorption onto suspended or bottom sediments, and microbial degradation⁶. The aim of the present study was to optimise some physicochemical conditions such as pH, temperature and carbon nitrogen ratio for the growth and degradation of most persistent mono-chlorophenol, *m*-chlorophenol by *Pseudomonas putida* CP1.

P. putida CP1 was obtained from Dr. Favio Fava, University of Bologna, Italy. It was maintained on agar medium containing mchlorophenol (Aldrich Chemical Co, UK) at 4°C for and subcutured monthly. Inoculum was prepared by growing the bacterium overnight in nutrient broth, harvesting cells by centrifuged at 5,000 rpm for 10 min, washing twice with 0.01 M sodium phosphate buffer and re-suspending in distilled water. Culture suspension (5 ml) was used to inoculate 95 ml sterile minimal medium⁷ containing m-chlorophenol in 250-ml conical flasks. The minimal medium with an initial pH 7.0 contained 4.36 g/l K₂HPO₄, 3.45 g/l NaH₂PO₄, $1.26 \text{ g/l}(\text{NH}_{A})_2 \text{SO}_4, 0.912 \text{ g/l} \text{MgSO}_4.7\text{H}_2\text{O} \text{ and } 1 \text{ ml/l} \text{ trace solution}.$ Trace salts solution contained (per 100 ml) 4.77 g CaCl₂.2H₂O, 0.37 g FeSO₄.7H₂O, 0.37 g CoCl₂.6H₂O, 0.10 g MnCl₂.4H₂O and 0.02 g Na₂MoO₄.2H₂O. After inoculation, flasks were incubated in an orbital shaker at 150 rpm at 30°C. Samples were aseptically removed at regular intervals and analysed for growth, pH, chlorophenol removal, COD and for reducing sugar where appropriate.

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Growth of P. putida CP1 was monitored by turbidity measurement. For the determination of pH dependency on *m*-chlorophenol removal by the bacterium, the pH of the minimal medium was adjusted to various pH values (5.0-8.0), and then m-chlorophenol (200 ppm) was added and they were incubated at 30°C and 150 rpm in an orbital shaker for desired period. To optimise the cultivation temperature, the bacterium was grown in the minimal medium with an initial pH 6.5 at different temperatures (20-37°C) under shaking (150 rpm in an orbital shaker. After incubation cells were separated from culture medium by centrifugation. Clear supernatant was used for the assay of the *m*-chlorophenol. pH and biomass level were also determined at different intervals. *m*-Chlorophenol concentrations were determined by using the 4-aminoantipyrene colorimetric method essentially as described in APHA⁸. Fructose concentration was determined by the dinitrosalicylate (DNS) method⁹. Chemical oxygen demand (COD) was determined using a modification of the method essentially as described in APHA8.

The initial pH of the culture media has a profound effect on the growth and removal of *m*-chlorophenol by *P. putida* CP1 (Figure 1). The maximum degradation of *m*-chlorophenol observed at pH between 6.5 and 7.0, which corresponded to complete removal of 200 ppm *m*-chlorophenol within 190 h. The organism could not grow or remove *m*-chlorophenol at pH 5.0 or above pH 7.0. At pH 6.0, only about 50% chlorophenol degradation was observed in 48 h. Khan *et al.*¹⁰ reported an optimum degradation of 4-aminophenol by *Pseudomonas* species ST-4 at pH 8.0.

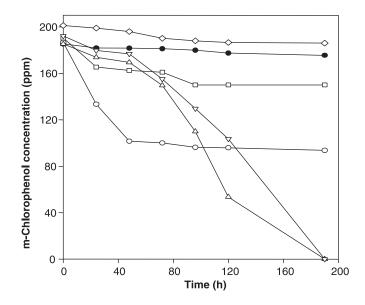


Figure 1. Removal of m-chlorophenol Pseudomonas putida CP1 at different pH containing minimal medium.

 $\bullet = pH \, 5; \ \Box = pH \, 5.5; \ \circ = pH \, 6; \ \vartriangle = pH \, 6.5; \ \nabla = 7; \ \diamond = pH \, 7.5$

Effect of cultivation temperature on *m*-chlorophenol removal by *P. putida* CP1 is shown in Figure 2. The bacterium showed the

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temperature optimum for removal *m*-chlorophenol at 30°C. It was unable to completely remove *m*-chlorophenol at a concentration of 200 ppm at temperatures above and below 30°C. The 4-aminophenol degradation by *Pseudomonas* species ST-4 had been reported at 30°C by Khan *et al.*¹⁰.

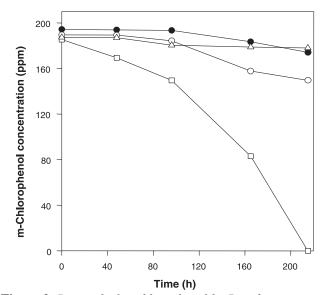


Figure 2. Removal of m-chlorophenol by Pseudomonas putida CP1 at different temperature at pH 6.5 containing minimal medium

• = $20^{\circ}C$; • = $25^{\circ}C$; \Box = $30^{\circ}C$; Δ = $37^{\circ}C$

It was found in our previous studies¹¹⁻¹² that fructose and yeast extract had positive influence on *m*-chlorophenol degradation by P. putida CP1. In this study, the influence of various combinations of fructose and yeast extract on degradation of m-chlorophenol by P. putida CP1 was studied. It was observed that more COD was removed in the presence of yeast extract plus fructose compare to either yeast extract or fructose alone (Figure 3). Complete removal of *m*-chlorophenol was achieved within 22 h in presence yeast extract plus fructose. The mchlorophenol removal rate was 1.03 mg/l/h in the absence of fructose and yeast extract (Table 1). On the other hand, the rate was 7.467, 7.529, 3.01, 4.60, 9.31 mg/l/h in the presence of 0.2% fructose, 0.5% fructose, 0.1% yeast extract, 0.2% yeast extract, 0.1% yeast extract plus 0.2% fructose and 0.2 yeast extract plus 0.5% fructose respectively. Conventional carbon sources may have some other effects, as they may provide the reducing power for the degradation of recalcitrant organic compounds¹³ or in some cases act as inducing agents for biodegradative enzymes¹⁴. Carbon : nitrogen ratio (3:1) was found best for the complete removal of COD by P. putida CP1, which correlated the optimum C:N for bacteria¹⁵. It could be concluded from the present study that P. putida CP1 could effectively remove COD and degrade *m*-chlorophenol in the presence of both carbon (fructose) and nitrogen (yeast extract) sources.

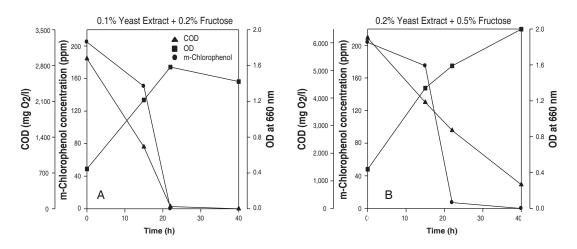


Figure 3. Removal of m-chlorophenol (200 ppm) by Pseudomonas putida CP1 in the presence of (A) 0.1% yeast extract plus 0.2% fructose and (B) 0.2% yeast extract plus 0.5% fructose. COD = Chemical oxygen demand; OD = Optical density

Table 1. Influence of carbon : nitrogen (C:N) ratio on rate of degradation of m-chlorophenol (200 ppm) by Pseudomonas putida CP1.

Treatment	C:N ratio	Δ pH	Lag-cp (h)	Chlorophenol removal rate (mg/l/h)	Fructose removal rate (mg/l/h)
cp + 0.2% F	3.48	- 0.24	15	7.47	44.45
cp + 0.5% F	8.06	- 0.55	15	7.53	126.16
cp + 0.1% YE	0.804	- 0.02	6	3.01	-
cp + 0.2% YE	1.013	0.08	6	4.6	-
$cp + 0.1\% \ YE + 0.2\% \ F$	2.96	- 0.16	-	9.31	87.68
$cp + 0.2\% \ YE + 0.5\% \ F$	5.17	- 0.38	-	8.93	119.2

Lag-cp = Lag period for m-chlorophenol degradation; cp = m-chlorophenol; F = fructose; YE = yeast extract

References

- Fetzner S & Lingens F. 1994. Bacterial dehalogenases: Biochemistry, genetics, and biotechnological applications. *Microbiol Rev.* 58: 641-685.
- Puhakka JA & Melin ES. 1996. Bioremediation of chlorinated phenols. In *Bioremediation: Principles and Applications* (Crawford RL & Crawford DL eds), pp 254-299. Biotechnology Research Series: 6. Cambridge University Press, Cambridge.
- Tabak HH, Chamber CW & Kabler DW. 1964. Microbial metabolism of aromatic compounds. I. Decomposition of phenolic compounds and aromatic hydrocarbons by phenol-adapted bacteria. *J Bacteriol.* 87: 910-919.
- Alexander M & Aleem MIH. 1961. Effect of chemical structure on microbial decomposition of aromatic herbicides. *J Agric Food Chem.* 9: 44-47.
- Alexander M & Lustigman BK. 1966. Effect of chemical structure on microbial decomposition of substituted benzene. *J Agric Food Chem.* 14: 410-413.
- Hale DD, Reineke W & Wiegel J. 1994. Chlorophenol degradation. In *Biological Degradation and Bioremediation of Toxic Chemicals* (Chaudhry GR ed), pp 74-91. Chapman & Hall, London.
- Goulding C, Giller CJ & Bolton E. 1988. Biodegradation of substituted benzenes. J Appl Bacteriol. 65: 1-5.

- APHA. 1998. Aggregate organic constituents. In *Standard Methods* for the Examination of Water and Wastewater (Greenberg AE, Clesceri LS & Eaton AD eds), 20th edn. American Public Health Association, Washinton DC.
- 9. Miller GL 1959. Use of dinitrosalicylic acid reagent for the determination of reducing sugar. *Anal Chem.* **31**: 426-428.
- Khan SA, Hamayun M & Ahmed S. 2005. Optimumisation of some physicochemical conditions for the degradation of 4-aminophenol by newly isolated *Pseudomonas* species ST-4. *Bangladesh J Microbiol.* 22: 109-112.
- Fakhruddin ANM. 2005. Effect of readily metabolisable substrates on the degradation of 3-chlorophenol by *Pseudomonas putida* CP1. *Nucl Sci Appl.* 14: 112-118.
- Fakhruddin ANM & Quilty B. 2005. The influence of glucose and fructose on the degradation of 2-chlorophenol by *Pseudomonas putida* CP1. World J Microbiol Biotechnol. 21: 1541-1548.
- Perkins PS, Komisar SJ, Puhakka JA & Ferguson JF. 1994. Effects of electron donors and inhibitors on reductive dechlorination of 2,4,6trichlorophenol. *Wat Res.* 28: 2101-2107.
- Chaudhuri BK & Wiesmann U 1995. Enhanced anaerobic degradation of benzene by enrichment of mixed microbial culture and optimisation of the culture medium. *Appl Microbiol Biotechnol.* 43: 178-187.
- Stanbury PF & Whitaker A. 1984. Media for industrial fermentations. In *Principles of Fermentation Technology*, 1st edn, pp 74-89. Pergamon Press, New York.